

1                   **Air Quality-Carbon-Water Synergies and Trade-offs**  
2                                   **in China's Natural Gas Industry**

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25 **Abstract**

26 **Both energy production and consumption can simultaneously affect regional air**  
27 **quality, local water stress, and the global climate. Identifying the air quality-carbon-**  
28 **water interactions due to both energy sources and end-uses is important for**  
29 **capturing potential co-benefits while avoiding unintended consequences when**  
30 **designing sustainable energy transition pathways. Here, we examine the air quality-**  
31 **carbon-water interdependencies of China’s six major natural gas sources and three**  
32 **end-use gas-for-coal substitution strategies in 2020. We find that replacing coal with**  
33 **gas sources other than coal-based synthetic natural gas (SNG) generally offer**  
34 **national air quality-carbon-water co-benefits. However, SNG achieves air quality**  
35 **benefits while increasing carbon emissions and water demand, particularly in**  
36 **regions already suffering from high per capita carbon emissions and severe water**  
37 **scarcity. Depending on end-uses, non-SNG gas-for-coal substitution results in**  
38 **enormous variations in air quality, carbon, and water improvements, with notable**  
39 **air quality-carbon synergies but air quality-water trade-offs. This indicates that**  
40 **more attention is needed to determine in which end-uses natural gas should be**  
41 **deployed to achieve desired environmental improvements. Assessing air quality-**  
42 **carbon-water impacts across local, regional and global administrative levels is**  
43 **crucial for designing and balancing the co-benefits of sustainable energy**  
44 **development and deployment policies at all scales.**

45

46 Most fossil energy production and combustion processes emit air pollutants and  
47 greenhouse gases (GHGs) and also consume significant quantities of freshwater<sup>1,2</sup>.

48 Depending on differences in fuel types, burning conditions, cooling techniques, and  
49 existing local environmental stress, energy source choices and end-uses can lead to  
50 substantial variations in the resulting air quality, climate, and water impacts<sup>4,9</sup>. Previous  
51 studies have concentrated on one or in some cases two, specific environmental impacts in  
52 the energy industry<sup>5,11</sup>. Very few analyses have evaluated the air quality-carbon-water  
53 interrelationships of the energy sector<sup>12-13</sup>, and even fewer have analyzed the nexus from  
54 both supply and end-use perspectives<sup>14-16</sup>. Characterizing the interconnections of various  
55 environmental impacts resulting from energy source choices and end-use applications is  
56 critical in achieving air quality, carbon and water co-benefits while avoiding unintended  
57 side effects. Here we examine China's natural gas industry and systematically analyze the  
58 synergies and trade-offs among the air quality, carbon, and water impacts due to both  
59 natural gas source choices (from where natural gas originates) and deployment strategies  
60 (in which region and subsector natural gas is substituted for coal).

61 Similar to many emerging economies, China has been facing multiple environmental  
62 challenges including domestic air pollution, local water scarcity, and global climate  
63 change<sup>1,3,17</sup>. A coal-dominated energy structure (~64% of primary energy supply in 2015)<sup>18</sup>  
64 is partly responsible for all three environmental stresses<sup>1,3,19</sup>. Natural gas is the cleanest  
65 fossil fuel, with relatively low carbon intensity and lower cooling water requirements  
66 than coal in most end uses<sup>3, 6, 20</sup>. Primarily to tackle its severe air pollution and the  
67 associated human health impacts<sup>3,21</sup>, China has been actively promoting a coal-to-gas end-  
68 use energy transition<sup>2</sup>. Specifically, China plans to increase natural gas consumption from  
69 approximately 6% (~190 billion cubic meters, bcm) of national total primary energy  
70 consumption in 2015 to 10% (~360 bcm) in 2020<sup>17,23</sup>. Until recently, China's natural gas

71 supplies were primarily from domestic conventional gas production (~70%), imported  
72 liquefied natural gas (LNG) (~15%), and imported pipeline gas from Central Asian  
73 pipeline gas (~15%)<sup>24,25</sup>. To further increase gas supplies, China plans to develop domestic  
74 unconventional natural gas. For instance, China's latest government plans (issued in  
75 December 2016) aim to have an annual production of approximately 20 and 30 bcm of  
76 domestic coal-based synthetic natural gas (SNG) and shale gas, respectively, by 2020<sup>23,26</sup>.  
77 Meanwhile, China also plans to expand LNG annual import capacity by 38 bcm, as well  
78 as increasing pipeline gas from Russia and Central Asia by 38 and 30 bcm, respectively,  
79 by around 2020<sup>27,28</sup>.

80       Substituting conventional natural gas for coal is likely to bring multiple  
81 environmental benefits. However, the air quality, carbon, and water impacts, and their  
82 interactions at both aggregated and spatially resolved scales can vary depending on gas  
83 sources<sup>29-32</sup>. In addition, the magnitude and interdependencies of various environmental  
84 impacts of all gas sources can be affected by different end-use deployment strategies<sup>11,33</sup>.  
85 Earlier studies evaluated the air quality, carbon, and water impacts of the natural gas  
86 industry, with a focus on a specific gas source and on one (or in a few cases two)  
87 environmental impact(s)<sup>7,11,31,33-37</sup>. Few studies compared the lifecycle air pollutant or GHG  
88 emissions for SNG, LNG and conventional gas in the power sector<sup>30,32</sup>, or simultaneously  
89 calculated air pollutant emissions, GHG emissions, and water consumption for shale gas-  
90 fired electricity<sup>14</sup>. In this work, we integrate the analysis of various natural gas sources and  
91 end-uses to identify the underlying air quality-carbon-water synergies and trade-offs, as  
92 well as to understand the relative importance of gas source choices and end-uses in  
93 determining the environmental outcomes.

94 We use an integrated assessment approach in conjunction with lifecycle analysis to  
95 quantify net changes in China's air quality, carbon, and water impacts resulting from a  
96 fixed quantity (30 bcm, Supplementary Methods 1.1) of gas substituting for coal using  
97 each of China's six primary gas sources under three different deployment strategies  
98 (Supplementary Table 1). Essentially, we estimate changes in China's population-  
99 weighted air pollution concentrations, lifecycle GHG emissions, and water stress index  
100 (ranging from 0 to 1)<sup>38, 39</sup> weighted water consumption (referred hereafter as 'weighted  
101 water consumption') for each of 18 gas-for-coal substitution scenarios (Supplementary  
102 Tables 2-7). Comparing the multiple environmental impacts resulting from the  
103 deployment of various gas sources in different end-uses, we identify the multi-aspect  
104 environmental performance for each gas source and end-use combination, and  
105 characterize the resulting air quality-carbon-water interrelationships. Based on  
106 government and industrial plans<sup>11, 18, 23, 22, 40</sup>, Fig.1 shows, for each gas source, the spatial  
107 distribution of China's gas production and (or) the provinces in which each gas source  
108 will likely be consumed. For each gas source, we design three gas-for-coal end-use  
109 deployment strategies to reflect three different environmental priorities, including 1) air  
110 quality-focused substitution (AS), 2) carbon-focused substitution (CS), and 3) water-  
111 focused substitution (WS), respectively (Supplementary Table 1). We first estimate  
112 changes in air pollutant emissions, CO<sub>2</sub> emissions, and weighted water consumption  
113 resulting from end-use substitution of each gas source for coal under each deployment  
114 strategy. End-use gas substitution for coal results in an increase in natural gas demand  
115 and a decrease in coal demand. Both upstream natural gas and coal processes (i.e.,  
116 production, processing, transmission, and distribution) emit air pollutants and GHGs

117 (CO<sub>2eq</sub>, including both CO<sub>2</sub> and CH<sub>4</sub>) due to energy combustion and methane leakage<sup>31</sup>, and  
118 consume freshwater for dust suppression, coal washing, well drilling and other purposes<sup>41</sup>  
119 (Supplementary Table 8). Thus, we further quantify the corresponding environmental  
120 impacts due to increases in upstream gas processes and decreases in coal processes  
121 (Supplementary Table 9). Integrating upstream and end-use processes, we estimate net  
122 changes in air pollutant emissions, GHG emissions, and weighted water consumption for  
123 18 combinations of gas source choices and deployment strategies. Using the Weather  
124 Research and Forecasting model coupled with chemistry (WRF-Chem v3.6), we further  
125 simulate the resulting changes in PM<sub>2.5</sub> surface concentrations for each gas source under  
126 each deployment strategy and calculate the resulting changes in population-weighted  
127 PM<sub>2.5</sub> concentrations.

128

## 129 **Results**

### 130 **Aggregated air-carbon-water impacts from natural gas sources**

131 We estimate net changes in lifecycle air pollutant emissions, GHG emissions, and  
132 weighted water consumption for upstream and downstream stages of coal substitution by  
133 various gas sources in 2020. Fig.2 shows results for the air quality-focused substitution,  
134 which is the most likely scenario given China's current focus on improving air quality.  
135 We observe striking air quality-carbon/water trade-offs with SNG, but generally air  
136 quality-carbon-water co-benefits for all other gas sources.

137 As shown in Fig.2, end-use gas substitution for coal dominates net reductions in air  
138 pollutant emissions regardless of gas source. Despite net air pollutant emission reductions  
139 for all gas sources, SNG upstream gas processes lead to substantial net increases in

140 lifecycle GHG emissions and weighted water consumption. Upstream SNG processes  
141 emit roughly 4-7 times more CO<sub>2eq</sub> than other gas sources. Consequently, SNG  
142 substitution for coal increases 2020 lifecycle CO<sub>2eq</sub> emissions by ~20 (40) megatonnes  
143 (Mt) under the 100-year (20-year) global warming potential (GWP). However, depending  
144 on the gas source, substituting coal with the same amount of other gas sources leads to  
145 approximately 60 to 120 (70 to 140) Mt of CO<sub>2eq</sub> emission reductions under GWP<sub>100</sub>  
146 (GWP<sub>20</sub>) assuming a mean methane leakage rate. This is consistent with earlier findings  
147 that SNG has substantially higher lifecycle GHG emissions than other gas sources when  
148 used for electricity generation<sup>10,12</sup>. Similarly, weighted water consumption from upstream  
149 SNG processes is roughly 20-190 times greater than other gas sources, varying depending  
150 on which gas source is compared. As a result, SNG leads to an increase of ~200 million  
151 cubic meters (Mm<sup>3</sup>) of lifecycle weighted water consumption in 2020, while other gas  
152 sources result in ~20 to 60 Mm<sup>3</sup> of net reductions. In comparison, water consumption due  
153 to upstream SNG processes (~290 Mm<sup>3</sup>) is ~10-30 times higher than other gas sources  
154 (~10-23 Mm<sup>3</sup>) (Supplementary Fig.1). Actually, increased water consumption due to  
155 SNG projects alone can require ~10% and ~5% of total industrial water consumption in  
156 Xinjiang and Inner Mongolia, respectively. Differences in actual water consumption  
157 (~10-30 times) are significantly smaller than differences in weighted water consumption  
158 (~20-190 times), indicating that SNG production generally occurs in locations that are  
159 comparatively more water scarce than other gas producing regions. We find gas source  
160 choice matters for national carbon and water concerns, primarily because SNG results in  
161 substantial net carbon and water penalties while having similar air quality benefits as  
162 other gas sources.

163 Other than SNG, all gas sources, when substituted for coal, bring net reductions in  
164 lifecycle air pollutant emissions, weighted water consumption, and GHG emissions  
165 (assuming a mean methane leakage rate). GHG emissions from upstream gas processes  
166 (except for SNG) are largely offset by decreases in GHG emissions due to less coal  
167 production. This is partly because China's coal industry has high GHG emission  
168 intensities due to substantial underground coal mining associated with high methane  
169 emissions and a low methane recovery rate<sup>41</sup>. However, without proper methane leakage  
170 control from the natural gas industry, coal substitution with gas sources other than SNG,  
171 particularly with shale gas, can also result in net increases in lifecycle GHG emissions  
172 under both GWPs (Fig.2, and Supplementary Figures 2 and 3). Also, our estimated  
173 upstream weighted water consumption from shale gas processes is relatively small  
174 although it consumes twice as much water as upstream conventional gas processes  
175 (Supplementary Table 9). This is because China's existing shale gas development is  
176 mainly concentrated in water-abundant Sichuan basin, the location of roughly half of  
177 China's total shale gas resources<sup>42</sup>. Nevertheless, as a quarter of China's shale gas  
178 resources are located in northern water scarce regions<sup>42</sup>, further geographic expansion of  
179 shale gas development would likely worsen water stress there.

180

### 181 **Spatial air-carbon-water impacts from natural gas sources**

182 In addition to evaluating the aggregated environmental impacts, we also explore the  
183 spatial characteristics of China's air quality-carbon-water nexus to characterize the  
184 unintended redistributive effects. Fig.3 shows the 2020 spatial distribution of net changes  
185 in SO<sub>2</sub> emissions, simulated PM<sub>2.5</sub> surface concentrations, GHG emissions, and weighted



186 water consumption within mainland China for each gas source substituting for coal under  
187 the air quality-focused substitution (AS). At the regional level, we find that all gas  
188 sources generally bring air quality-carbon-water co-benefits in developed eastern China.  
189 However, although promoting SNG can help alleviate the severe air pollution and  
190 associated human health impacts in populated eastern China (currently a major objective  
191 in China), it results in substantial carbon-water losses in northwestern China, indicating a  
192 negative spillover effect of China's air quality improvement policies.

193 As shown in Fig.3, all gas sources, via substituting for coal, generally bring net  
194 reductions in SO<sub>2</sub> emissions and PM<sub>2.5</sub> surface concentrations in well-developed eastern  
195 China, though there are slight increases in northwestern provinces primarily due to SNG  
196 or conventional gas production. For each gas source, the largest PM<sub>2.5</sub> concentration  
197 reductions are primarily concentrated in regions where substitution occurs. Although  
198 SNG results in only slight increases in PM<sub>2.5</sub> surface concentrations in northwestern  
199 provinces, it leads to substantial increases in GHG emissions and weighted water  
200 consumption in northwestern SNG producing provinces. Notably, these regions also  
201 suffer from severe water scarcity and have high per capita carbon emissions, due to their  
202 coal-dominated energy mix and substantial export of electricity to eastern China, such as  
203 Beijing and Tianjin<sup>20-21</sup>. Similar negative spillovers are also observed from the carbon (CS)  
204 and water focused (WS) substitution (Supplementary Figs. 4 and 5).

205

### 206 **Air-carbon-water impacts from source choices and end-uses**

207 Besides gas source choices, sectoral and regional gas deployment strategies can also lead  
208 to large variations in net air quality, carbon and water impacts of gas substitution for coal.

209 To fully capture the synergies and trade-offs in the natural gas industry, here we integrate  
210 six major gas source choices and three end-use deployment strategies that affect the air  
211 quality-carbon-water interdependencies.

212 We find that gas source choice is the most critical factor in shaping the air quality-  
213 carbon-water nexus of the natural gas system, primarily because SNG clearly stands out.  
214 Unlike other gas sources, SNG worsens carbon emissions and water stress regardless of  
215 end-use deployment strategies (Fig.4). That is, SNG causes net increases in GHG  
216 emissions and weighted water consumption even under the deployment strategies that  
217 aim to achieve the largest reductions in GHG emissions (CS) or weighted water  
218 consumption (WS), respectively (Supplementary Figs. 2 and 3).

219 However, for gas sources other than SNG, end-uses determine the magnitude of the  
220 air quality, carbon, and water impacts. We find that within the same gas source (except  
221 for SNG), different *deployment* strategies can result in more than 10-50 times differences  
222 in China's population-weighted (P-W) PM<sub>2.5</sub> surface concentration reductions (Fig.4,  
223 Supplementary Methods 1.3). Similarly, different *deployment* strategies lead to  
224 approximately 1.5-1.6 times and 2-9 times variations in lifecycle GHG emissions (GWP<sub>20</sub>)  
225 and China's weighted water consumption reductions, respectively (Fig.4). In comparison,  
226 different gas *sources* lead to only roughly 1-4 times, 1-1.9 times, and 1-3 times  
227 differences, respectively, within the same deployment strategy. Thus, gas deployment  
228 strategies play a more critical role than gas source choices in determining the  
229 environmental impacts of non-SNG natural gas substitution for coal, particularly on air  
230 quality.

231        Additionally, gas deployment strategies are the key to determining the air quality-  
232 carbon-water interconnections for gas sources other than SNG. We note substantial air  
233 quality-carbon co-benefits but air quality-water trade-offs due to end-use gas substitution  
234 for coal. Specifically, depending on the gas source, AS leads to  $\sim 0.6$  to  $1.8 \mu\text{g}/\text{m}^3$  annual  
235 average P-W  $\text{PM}_{2.5}$  surface concentration reductions, but merely  $\sim 20$  to  $60 \text{ Mm}^3$  of  
236 weighted water consumption reductions within China in 2020. In comparison, varying on  
237 the gas source, WS results in  $\sim -0.004$  to  $0.07 \mu\text{g}/\text{m}^3$  P-W  $\text{PM}_{2.5}$  mean surface  
238 concentration reductions, but  $\sim 90$  to  $200 \text{ Mm}^3$  of weighted water consumption reductions.  
239 That is, for the same gas source, AS results in over an order of magnitude greater P-W  
240  $\text{PM}_{2.5}$  surface concentration reductions than WS. However, WS results in approximately 2-  
241 8 times as much weighted water consumption reductions as AS. In comparison, CS  
242 generally leads to similar levels of air quality, water stress and carbon emission changes  
243 as AS. We find CS actually results in slightly higher net reductions in P-W  $\text{PM}_{2.5}$  surface  
244 concentrations than AS. This is mainly because the least efficient coal combustion  
245 happens to be the dirtiest, indicating potential air-carbon co-benefits. Thus, when natural  
246 gas is deployed primarily to improve air quality, China's current top priority, it will in  
247 most cases bring substantial carbon reduction co-benefits, but have negligible water  
248 benefits. However, if natural gas were to be allocated mainly to address water scarcity  
249 concerns, in most cases it would only slightly improve air quality, though it would still  
250 bring notable carbon reductions. Therefore, there is a fundamental air quality-water trade-  
251 off due to end-use gas-for-coal substitution.

252        To put our estimated environmental impacts in perspective, we also show the percent  
253 changes for each impact resulting from each gas source and end-use choice. As shown in

254 Supplementary Fig.6, 30 bcm of natural gas, by substituting ~1.5% to 2.2% of total coal  
255 consumption, can lead to approximately -4.8% to 0.01%, -1.8% to 0.9%, and -0.07% to  
256 0.08% net changes in China's P-W PM<sub>2.5</sub> surface concentrations, lifecycle GHG emissions  
257 (GWP<sub>20</sub>), and China's weighted water consumption in 2020. Our estimated percent  
258 contributions may be scaled up or down depending on actual increases in natural gas  
259 supplies. However, the relative trends across gas sources and end-uses in affecting  
260 various environmental impacts can be illustrative. Supplementary Fig.6 again  
261 demonstrates the determining role of SNG in causing the air quality-carbon/water trade-  
262 offs, as well as substantial variations in the resulting environmental impacts (air quality in  
263 particular), primarily due to gas substitution for coal in various end-use applications.

264

## 265 **Discussion**

266 An energy transition away from fossil fuels and towards an air-, carbon-, and water-  
267 friendly future at local, national and global scales is a critical component of sustainable  
268 development. A transition from coal-to-gas is taking place as renewable energy comes  
269 into wider use. Our study demonstrates that with careful natural gas source choices and  
270 end-use designs, switching from coal to natural gas can bring air quality, carbon, and  
271 water co-benefits, though with notable air quality-water trade-offs in the magnitude of the  
272 resulting improvements. However, gas source choices can be a determining factor in  
273 changing this picture due to coal-based synthetic natural gas. Upstream SNG processes,  
274 particularly SNG production, significantly increase both water stress and carbon intensity  
275 in regions (northwestern China) already suffering from severe water scarcity and high per  
276 capita carbon intensity<sup>30, 38</sup>. Therefore, although end-use SNG substitution for coal reduces

277 CO<sub>2</sub> emissions and often reduces water consumption as well, SNG not only leads to an  
278 increase in lifecycle carbon and water consumption for China as a whole, but also  
279 exacerbates existing environmental injustice caused by energy export to eastern China<sup>29</sup>.  
280 Our results clearly show a negative spillover effect of China's "Clean Air Act", the focus  
281 on improving air quality in the well-developed eastern provinces may increase CO<sub>2</sub>  
282 emissions and water stress in the less-developed northwestern provinces when  
283 substituting SNG for coal. Earlier studies identified SNG as a good candidate for  
284 conducting carbon capture and storage (CCS), as CO<sub>2</sub> emitted during SNG production is  
285 of high partial pressure and high purity<sup>11, 30</sup>. Assuming ~90% CO<sub>2</sub> removal efficiency  
286 during SNG production<sup>11</sup>, applying CCS with SNG could reduce CO<sub>2</sub> emissions by ~110  
287 Mt, resulting in net GHG reductions from a coal-to-SNG switch. However, development  
288 of CCS will further increase water demand in northwestern regions due to water  
289 consumption for CO<sub>2</sub> scrubbers and parasitic loads<sup>44, 45</sup>. Thus, though CCS can make SNG a  
290 more attractive energy choice from the air quality and carbon perspectives, it exacerbates  
291 existing water stress, particularly in northwestern provinces. As energy infrastructure  
292 typically operates for multiple decades, our findings indicate the need to identify the air  
293 quality-carbon-water interconnections before making large-scale energy investments to  
294 avoid unintended side effects at both regional and global levels.

295 For coal substitution with gas sources other than SNG, we find that end-use gas  
296 deployment usually plays a far more important role than gas source choices in  
297 determining the magnitude of resulting local air pollution and water stress alleviation, as  
298 well as carbon mitigation in most cases. Existing discussions have largely focused on  
299 clean energy source choices<sup>46</sup>. However, this study shows that more attention should be

300 placed on designing clean energy deployment strategies, as end-use choices can  
301 sometimes result in variations of over an order of magnitude in net environmental  
302 impacts.

303 Our study also illustrates notable air quality-water trade-offs due to end-use sectoral  
304 and regional natural gas deployment. This trade-off results from the sectoral differences  
305 affecting environmental impacts and the geographic mismatch between regions of high  
306 air pollution and high water stress (Supplementary Fig.7). Particularly, under WS, natural  
307 gas is primarily distributed to the power sector to substitute for coal (Supplementary Fig.  
308 8a). This allocation can significantly reduce water consumption but brings only small air  
309 quality benefits due to widely employed end-of-pipe control technologies in coal-fired  
310 power plants<sup>u</sup>. Conversely, when natural gas substitution for coal primarily occurs in the  
311 residential sector where it results in the largest reductions in air pollution emissions<sup>u</sup>, it  
312 does not reduce water consumption. Also, when more gas is allocated to highly water-  
313 stressed provinces under WS, it does not necessarily bring large reductions in air  
314 pollutant emissions because regions with high water stress and severe air pollution do not  
315 closely overlap (Supplementary Fig.8b). Notably, the inherent air quality-water trade-off  
316 identified here not only exists for coal substitution with natural gas, but also for coal  
317 substitution with renewables. For instance, displacing coal-fired power plants equipped  
318 with end-of-pipe controls with wind power will bring water savings but less pronounced  
319 air quality benefits. Thus, additional action in the residential and industrial sectors is  
320 necessary to achieve desired air quality improvements. Also, the trade-offs we identified  
321 may exist in other countries where regions of high air pollution differ from those with  
322 high water stress (i.e., India). Thus, we highlight the need for careful coordination of

323 energy and environmental policies to simultaneously and significantly address air quality,  
324 climate, and water concerns.

325       Additionally, both the air quality-carbon/water trade-offs due to energy source  
326 choices and the air quality-water trade-offs due to energy end-uses identified here  
327 highlight a conflict in decision making at the local, national, and global scales. Such  
328 conflicts widely exist across countries facing multiple environmental and energy  
329 challenges. Given the complexity of the air quality-carbon-water interactions at different  
330 administrative scales, there is no single optimal scenario that can outperform others in all  
331 three aspects. However, we do find that the air quality- and carbon-focused substitution  
332 scenarios with conventional and imported pipeline natural gas usually bring the most air  
333 quality and carbon benefits, and thus help to address China's current primary concerns.  
334 Nevertheless, additional efforts are needed to achieve overall environmental  
335 improvements. For instance, limiting or curtailing the utilization of energy sources that  
336 result in substantial trade-offs (i.e., SNG) or planning a combination of technologies that  
337 compensate for potential trade-offs (i.e., gas-for-coal substitution in the residential sector  
338 coupled with dry-cooling technology in the power sector) may reduce trade-offs at  
339 varying administrative scales. In addition, from the perspective of policy implementation,  
340 it is important to consider the economic costs of gas source and end-use options. For  
341 instance, unconventional natural gas generally costs more than conventional and  
342 imported pipeline gas due to smaller scale production and immature technology for  
343 unconventional gas<sup>23,31</sup>. This further disfavors SNG, whose future production scale should  
344 be limited due to carbon and water concerns. Also, although deploying natural gas in the  
345 residential sector (mainly under the AS and CS scenarios) brings the most air quality and

346 carbon benefits, it is usually very costly due to the need to install expensive last-meter  
347 distribution pipelines<sup>u</sup>. Thus, government subsidies for residential sector gas  
348 infrastructure is needed to facilitate an end-use coal-to-gas conversion for residents<sup>s</sup>.  
349 Further analysis of the regional variations and dynamic changes in economic costs are  
350 needed to better evaluate the feasibility of different gas source choices and end-use  
351 designs at finer resolution.

352 The absolute environmental impacts we estimate may vary depending on actual  
353 increases in natural gas supplies, actual baseline energy consumption, the penetration and  
354 removal rates of sub-sectoral end-of-pipe control technologies, and the non-linearity of  
355 atmospheric chemistry. This non-linearity may also change the order of air quality  
356 benefits for scenarios, especially when existing differences across scenarios are small.  
357 Due to the large computational resources required to simulate air pollution concentrations  
358 for all gas source and end-use combinations, we choose a representative additional gas  
359 supply and a widely used emission scenario as the base case in this study (Methodology  
360 and Supplementary Methods 1.1)<sup>u,42,48</sup>. We compare the air quality-carbon-water impacts  
361 among various gas sources and end-use designs, all of which have the same baseline and  
362 the same quantity of additional gas supply. Thus, the underlying air quality-carbon-water  
363 synergies and trade-offs resulting from various gas sources and end-uses should remain  
364 the same.

365 Globally, there are large uncertainties in methane leakage rates from upstream natural  
366 gas processes. This can potentially make gas source choices a more important factor in  
367 affecting net carbon impacts than we identify here. Field measurements of methane  
368 leakage along the whole lifecycle chain of the natural gas industry both within and



369 outside China will improve understanding of the carbon impacts of China's natural gas  
370 industry and the relative importance of gas source choices and deployment strategies.

371 The air quality-carbon-water nexus discussed in this study focuses on China's natural  
372 gas industry. Due to its enormous economy and population, China's energy plans have  
373 important domestic as well as global implications for sustainable development. The  
374 framework described here, and its qualitative conclusions could be applied to other  
375 countries and regions as they design sustainable energy transition pathways.

376

### 377 **Methods**

378 This work uses an integrated assessment approach coupled with lifecycle analysis to evaluate the  
379 air quality-carbon-water nexus of China's natural gas industry. Our objective is to understand the  
380 differences in impacts resulting from various gas source choices and end-use deployment  
381 strategies (Supplementary Table 1).

382 We quantify the air quality impacts as changes in lifecycle air pollutant emissions and  
383 simulated PM<sub>2.5</sub> surface concentrations. Carbon impacts are calculated as changes in lifecycle  
384 greenhouse gas (GHG) emissions. Additionally, water impacts are represented by changes in  
385 water consumption weighted by water stress index (WSI) (WSI: the ratio of total annual  
386 freshwater withdrawal to hydrological availability, ranging from 0 to 1<sup>33,34</sup>). WSI weighted water  
387 consumption is calculated as actual regional water consumption × region-specific water stress  
388 index, referred hereafter as 'weighted water consumption'.

389 We use the ECLIPSE\_V5a\_CLE (Evaluating the Climate and Air Quality Impacts of Short-  
390 Lived Pollutants) emission scenario developed by the Greenhouse Gas and Air Pollution  
391 Interactions and Synergies (GAINS) model as our 2020 base case anthropogenic emissions  
392 input<sup>35</sup>. The ECLIPSE scenario is designed to reflect provincial energy policies and emission  
393 regulations in China's 12<sup>th</sup> five-year-plan<sup>36</sup>, and it provides detailed subsector technology

394 information, energy consumption data, and emissions of major air pollutants and CO<sub>2</sub> at China's  
395 provincial level. In addition, we integrate China's provincial-level cooling technology  
396 information from the World Electric Power Plants database (<https://www.platts.com/products/>)  
397 (Supplementary Table 2) with end-use technology data (i.e., power plant technologies) provided  
398 by the GAINS model (<http://gains.iiasa.ac.at/models/>) to evaluate water impacts (Supplementary  
399 Methods).

400 We focus on China's six major natural gas sources, including domestic conventional natural  
401 gas, domestic coal-based synthetic natural gas (SNG), domestic shale gas, imported liquefied  
402 natural gas (LNG), imported pipeline gas from Russia, and imported pipeline gas from Central  
403 Asia. For each gas source, the regional deployment is determined by governmental and industrial  
404 plans as shown in Fig.1 and Supplementary Table 4. At the sectoral level, we consider natural gas  
405 substitution for coal in three major sectors: industry, residential, and power<sup>u</sup>.

406 Based on three possible environmental priorities, for each gas source, we design three end-use  
407 deployment strategies for the gas-for-coal substitution for each gas source. 1) Air quality-focused  
408 substitution (AS), designed to achieve the largest reductions in SO<sub>2</sub> emissions; 2) Carbon-focused  
409 substitution (CS), designed to achieve the largest reductions in CO<sub>2</sub> emissions; and 3) Water-  
410 focused substitution (WS), designed to achieve the largest reductions in weighted water  
411 consumption (Supplementary Table 1).

412 To uniformly compare impacts of gas-for-coal substitution, for each combination of gas  
413 source and end-use, we assume an additional gas supply of 30 bcm above the baseline to replace  
414 coal. This is roughly the quantity of China's currently planned increases for each gas source  
415 around 2020<sup>21, 22, 23</sup> (Supplementary Methods 1.1). We then estimate the resulting changes in air  
416 pollutant emissions, CO<sub>2</sub> emissions, and weighted water consumption due to end-use gas  
417 substitution for coal (Supplementary Tables 2 and 3, Supplementary Fig. 7). Additional gas  
418 supply leads to an increase in upstream natural gas production and a decrease in upstream coal  
419 production. This results in emission and water consumption changes from upstream natural gas

420 and coal processes (i.e., production, processing, transmission, and distribution) due to energy  
421 combustion, methane leakage, and water uses for coal washing, well drilling and so forth<sup>u,u</sup>  
422 (Supplementary Table 5). We quantify changes in upstream emissions primarily using stage-level  
423 energy consumption data and methane leakage rates summarized in Qin et al. (2017)<sup>u</sup> (refer to  
424 Supplementary Methods 1.2 for details), and country-specific emission factors from the GAINS  
425 model (Supplementary Fig. 10). We calculate changes in upstream weighted water consumption  
426 using the same energy consumption data and methane leakage rates (Qin et al., 2017<sup>u</sup>), fuel-  
427 specific water consumption rates (Supplementary Table 8), and regional water stress indexes.  
428 WSI are provided by Feng et al. (2014)<sup>u</sup> (provincial-level WSI for regions within China) and  
429 Pfister et al. (2009)<sup>u</sup> (country-level WSI for regions outside China). We assume upstream  
430 emissions and water consumption for gas processes occur in places where natural gas is  
431 produced. The spatial distribution of reduced emissions and water consumption due to avoided  
432 upstream coal processes are identified according to where end-use coal reduction occurs and the  
433 corresponding source-receptor matrix of coal production and consumption<sup>u</sup>.

434 Combining both upstream and end-use processes, we estimate net changes in air pollutant  
435 emissions, GHG emissions, and weighted water consumption for each gas source under each gas-  
436 for-coal deployment strategy. We then use the Weather Research and Forecasting model coupled  
437 with Chemistry (WRF-Chem v3.6) to simulate resulting changes in 2020 annual average PM<sub>2.5</sub>  
438 surface concentrations. Method details are summarized in the Supplementary Methods.

439

#### 440 **Data availability**

441 Data used to perform this work can be found in the Supplementary Information. Any  
442 further data that support the findings of this study are available from the corresponding  
443 authors upon reasonable request.

444

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446 mauzeral@princeton.edu and **Yue Qin:** yq@princeton.edu or yqin8@uci.edu.

447

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455

#### 456 **Author Contributions**

457 Y.Q. and D.L.M. designed the study, Y.Q. performed the research, L.H.I, E.B., K.F.,  
458 F.W., and W.P. contributed data for analysis, Y.Q., L.H.I, E.B., and K.F. analyzed data,  
459 and Y.Q., D.L.M., and L.H.I wrote the paper.

460

#### 461 **Competing interests**

462 The authors declare no competing financial interests.

463 References

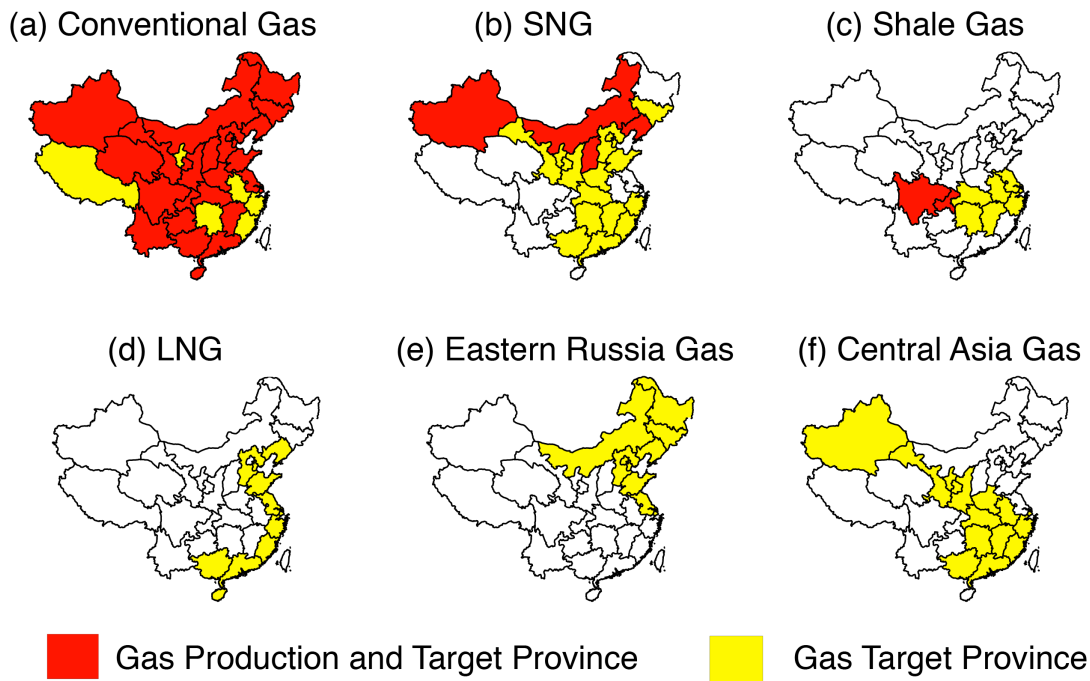
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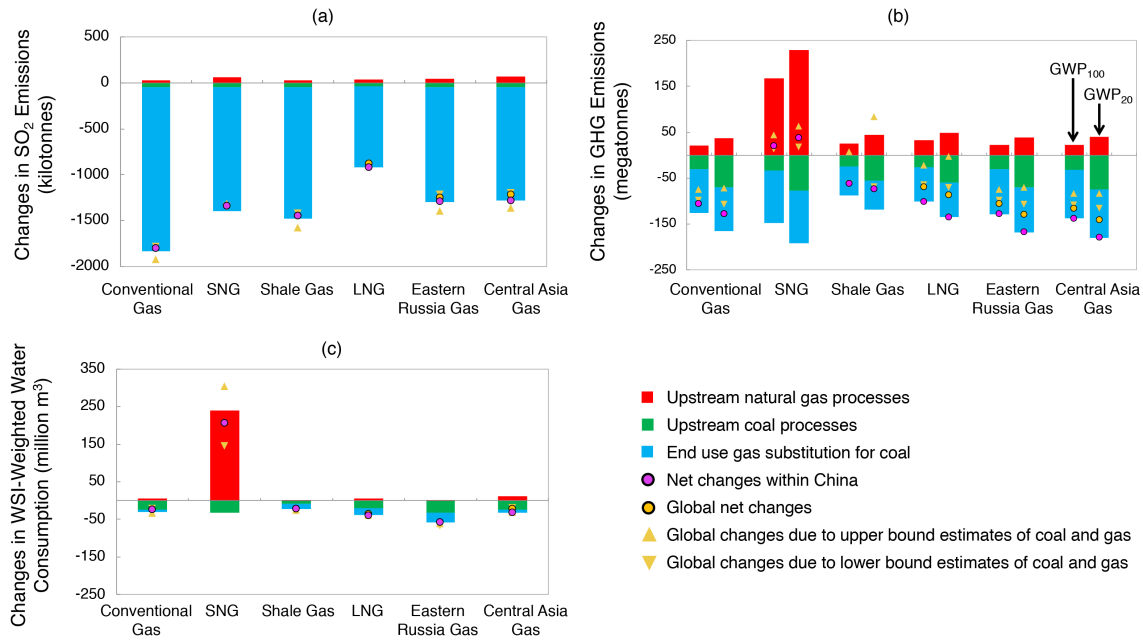
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649 Fig. 1 Production and target (potential consumption) regions for mainland China's six major  
 650 natural gas sources based on government and industrial plans for 2020<sup>11, 18, 21, 22, 40</sup>. Refer to  
 651 Supplementary Table 4 for details and the underlying assumptions about the spatial distribution  
 652 of gas production and consumption. Imported LNG is mainly produced in Qatar, Australia,  
 653 Indonesia, and Malaysia<sup>41</sup>. Imported Eastern Russia Gas and Central Asia Gas are produced in  
 654 Russia and Central Asian countries (Turkmenistan, Kazakhstan, and Uzbekistan), respectively<sup>22,28</sup>.

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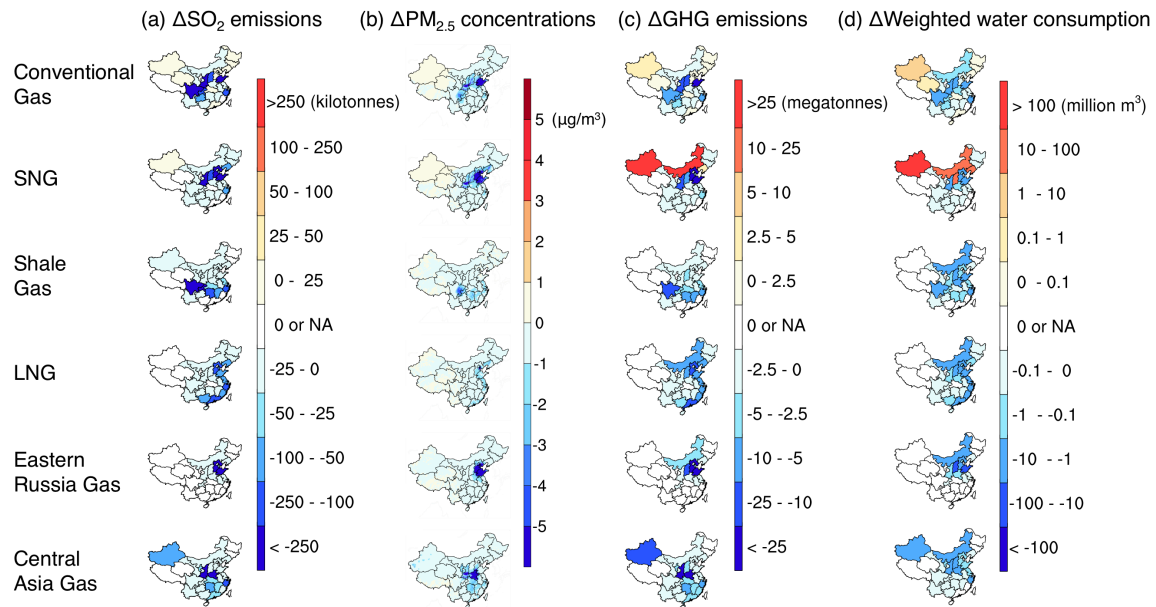


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659 Fig. 2 **Air quality-focused substitution (AS)**. Changes in upstream and downstream SO<sub>2</sub>  
 660 emissions, greenhouse gas emissions (CO<sub>2eq</sub>, including both CO<sub>2</sub> and CH<sub>4</sub>), and water stress index  
 661 (WSI) weighted water consumption for coal substitution by 30 bcm of gas from various sources  
 662 (Conventional gas, SNG, Shale gas, LNG, Eastern Russia gas, and Central Asia gas) in 2020. **Net**  
 663 **changes within China** are obtained by considering emission or weighted water consumption  
 664 changes occurring **within China's border** from upstream gas processes (upstream gas  
 665 production, processing, transmission, and distribution), upstream coal processes (upstream coal  
 666 production, processing, and transport), and end-use gas substitution for coal. **Global net changes**  
 667 represent the emission or weighted water consumption changes resulting from the differences  
 668 **within and outside of China**. Global changes due to upper and lower bound estimates (mainly  
 669 due to methane leakage rates) of coal and natural gas, respectively, are also shown in the graph.  
 670 Note that differences between the mean estimates of coal and the mean estimates of gas are not  
 671 necessarily the mean differences between coal and natural gas. Results for carbon-focused (CS)  
 672 and water-focused (WS) gas-for-coal substitution are shown in Supplementary Figs. 2 and 3.  
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677 **Fig. 3 Air quality-focused substitution (AS).** Mainland China's 2020 spatially resolved changes  
678 in air quality ( $\text{SO}_2$  emissions and population-weighted  $\text{PM}_{2.5}$  surface concentrations), carbon  
679 (lifecycle greenhouse gas emissions under  $\text{GWP}_{100}$ , including  $\text{CO}_2$  and  $\text{CH}_4$ , assuming mean  
680 methane leakage rates), and water (weighted water consumption) impacts of 30 bcm of gas from  
681 various sources substituting for coal. Changes in  $\text{SO}_2$  emissions, GHG emissions, and weighted  
682 water consumption are shown at the provincial level; changes in  $\text{PM}_{2.5}$  concentrations are shown at  
683 the grid level (27 km $\times$ 27 km). Results for carbon-focused (CS) and water-focused (WS) gas-for-  
684 coal substitution are shown in Supplementary Figs. 4 and 5.

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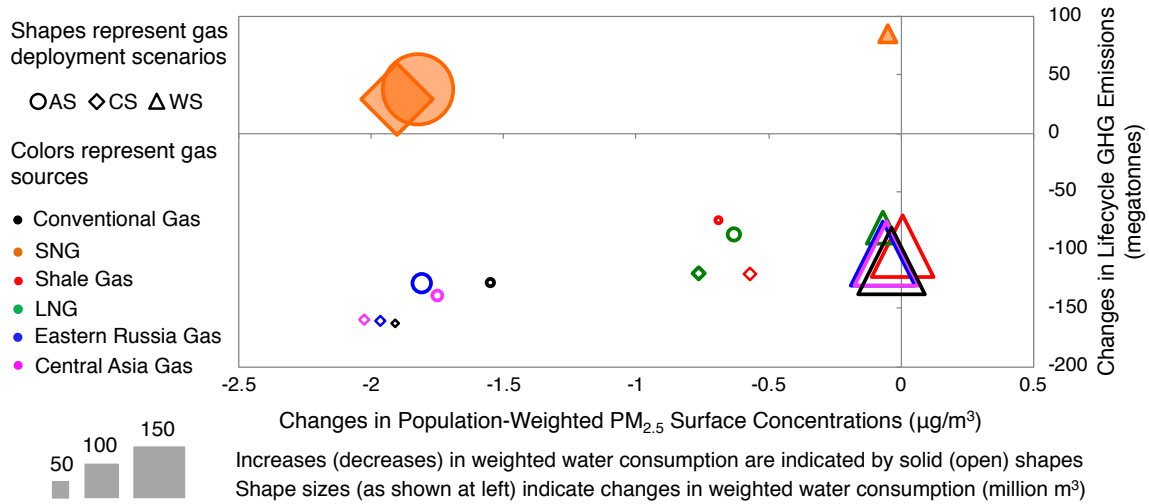
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699 Fig. 4 Comparison of net changes in air quality (China's population-weighted PM<sub>2.5</sub> surface  
700 concentrations), carbon (lifecycle greenhouse gas emissions under GWP<sub>20</sub>, assuming mean  
701 methane leakage rates), and water impacts (China's weighted water consumption) from  
702 substituting 30 bcm of gas from various sources for coal under three deployment strategies in  
703 2020. Negative changes in PM<sub>2.5</sub> surface concentrations, lifecycle GHG emissions, and weighted  
704 water consumption represent an improvement in air quality, carbon, and water impacts from gas  
705 substitution for coal, and vice versa. AS: air quality-focused substitution, CS: carbon-focused  
706 substitution, and WS: water-focused substitution. Detailed breakdowns are shown in  
707 Supplementary Fig. 9.

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